

1 **Title: Strong habitat and weak genetic effects shape the lifetime reproductive**
2 **success in a wild clownfish population**

3
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22 **Running title:** Habitat drives clownfish local success

Abstract

Lifetime reproductive success (LRS), the number of offspring an individual contributes to the next generation, is of fundamental importance in ecology and evolutionary biology. LRS may be influenced by environmental, maternal and additive genetic factors, and the relative contributions of each are critical in determining whether species can adapt to rapid environmental change. However, studies quantifying LRS across multiple generations in wild populations are extremely rare, and to date, non-existent for marine species. Here we use pedigrees of up to 5 generations resolved from a 10-year data-set for a wild orange clownfish population from Kimbe Island (PNG) to assess the contribution of every breeder to the local population. We quantified the additive genetic, maternal and environmental contributions to variation in LRS for the self-recruiting portion of the population using a genetic linear mixed model approach. We found that the habitat of the breeder, including the anemone species and geographic location, made the greatest contribution to LRS, explaining ~97% of the variation. There were low to negligible contributions of genetic (1.3%) and maternal factors (1.9%) equating with low heritability and evolvability. Our findings imply our population will be extremely susceptible to short-term, small-scale changes in habitat structure and may have limited capacity to adapt to these changes.

Keywords: multi-generational pedigree, additive genetic variation, maternal effects, environmental effects, adaptation, selection, heritability, evolvability.

42 INTRODUCTION

43 Darwinian fitness or lifetime reproductive success (LRS) – the number of successful offspring an
44 individual contributes to the next generation – is a critical variable underpinning ecological and
45 evolutionary responses to the environment (Clutton-Brock 1988; Roff 2002; Clutton-Brock &
46 Sheldon 2010; Hendry *et al.* 2018). LRS may be influenced by several factors including phenotypic
47 responses to different environmental parameters, parental effects, and additive genetic variance
48 (Kruuk *et al.* 2000; McCleery *et al.* 2004; McFarlane *et al.* 2014). Apportioning these contributions
49 is critical to understanding the potential for short-term ecological effects and longer-term
50 evolvability in response to rapid environmental change. If LRS is exclusively a phenotypic response
51 to the conditions experienced by individuals, populations will be severely impacted by rapid
52 environmental change and there is no prospect of adaptive microevolution. Maternal responses to
53 environmental conditions may reduce the effect of those changes on the next generation through
54 acclimation (Bernardo 1996; Mousseau & Fox 1998; Danchin *et al.* 2011; Donelson *et al.* 2017).
55 However, it is the additive genetic variation in LRS that determines the rate of adaptation of a
56 population to the environmental demand (Fisher 1930; Frank 1997; Lessard 1997; Crow 2002). To
57 date, few studies have distinguished the relative importance of environmental and genetic
58 contributions to LRS over multiple generations in wild populations, where genetic contributions to
59 fitness may be complex (Kruuk & Hadfield 2007). This situation is changing as long-term,
60 individual-based ecological studies in which multi-generational pedigrees have been established
61 provide the necessary longitudinal information to quantify the different components of LRS
62 (Pemberton 2008; Clutton-Brock & Sheldon 2010; Wolak *et al.* 2018; Cava *et al.* 2019). Such
63 studies are imperative as we seek to understand the ability of species to withstand or adapt to
64 accelerating climate change (Charmantier *et al.* 2008; Munday *et al.* 2017).

65

66 Intergenerational responses to selection are a product of the interplay between genetic and
67 environmental mechanisms that ultimately shape the genetic variation in fitness-related traits.

68 Environmentally driven mechanisms (*e.g.*, phenotypic plasticity, genetic assimilation) can facilitate
69 (Price *et al.* 2003; Ghalambor *et al.* 2007, Ledón-Rettig *et al.* 2014; Hoffmann & Merilä 1999;
70 Danchin *et al.* 2019) or constrain the microevolutionary response to selection (Kruuk *et al.* 2003;
71 Pujol *et al.* 2018). However, in the absence of genetic variation for LRS, these mechanisms will
72 likely have little effect on a negligible rate of adaptive evolution. In quantifying additive genetic
73 variation, it is important to distinguish between *heritability* and *evolvability* (Wheelwright *et al.*
74 2014; Cava *et al.* 2019). Narrow sense heritability, the additive genetic variance standardized by the
75 total phenotypic variance, is widely used as a measurement of the population evolutionary potential
76 to respond to selection (Falconer & Mackay 1996; Mousseau & Roff 1987; McCleery *et al.* 2004).
77 It is directly affected both by the additive genetic variation and magnitude of direct environmental
78 effects. Low heritability values can either reflect low additive genetic variance or large
79 environmental, or residual effects (Price & Schluter 1991; Houle 1992; Hansen *et al.* 2011).
80 Evolvability is the mean-standardized additive genetic variance (Houle 1992; Hansen *et al.* 2011).
81 This is the expected proportional change per generation in population mean fitness given a unit
82 strength of selection (Hendry *et al.* 2018). Evolvability is not affected by environmental or maternal
83 effects, which makes it a more appropriate metric in the comparison of evolutionary potential
84 between traits, populations and species. Heritability reveals whether the additive genetic variance
85 for LRS represents a non-trivial proportion of the total variance of LRS in the actual environmental
86 context of a given wild population. Together, evolvability and heritability inform us about how
87 much environmental change a wild population can withstand on the basis of its evolutionary
88 potential.

89

90 The few ($n = 15$) long-term, individual-based studies that have quantified additive genetic variation,
91 heritability and evolvability of LRS in wild populations have all focused on terrestrial species
92 (Postma 2014; Hendry *et al.* 2018; Table S1). These have largely confirmed that LRS have low
93 additive genetic variation and evolvability (~ 0.08), which nevertheless reflects some evolutionary

94 potential (Burt 1995; Hendry *et al.* 2018). Until recently, quantifying LRS in marine organisms with
95 a pelagic larval stage has been considered impossible because of the difficulties in following the
96 fate of offspring from one generation to the next. However, there is increasing evidence of some
97 degree of natal philopatry or self-recruitment in local marine populations (Jones *et al.* 1999;
98 Swearer *et al.* 1999; Swearer *et al.* 2002; Jones *et al.* 2009). The application of genetic parentage
99 analysis is making it possible to assign a significant proportion of successful offspring to their
100 parents (Planes *et al.* 2009; Jones 2015; Le Port *et al.* 2017; Mobley *et al.* 2019) and construct
101 multigenerational pedigrees (Salles *et al.* 2016a; Aykanat *et al.* 2014; Reed *et al.* 2019), at least for
102 the offspring that return to their natal population. In quantitative genetic studies of LRS, it is
103 impossible to measure the recruitment of dispersing juveniles at other locations. The regional
104 component of LRS, which would inform us on fitness variation beyond the local scale is impossible
105 to obtain. Measuring the local component of LRS in marine fish is an opportunity as in any other
106 species and estimates the relative contribution of local fish to the population self-recruitment and
107 replenishment.

108

109 For coral reef fishes, quantifying environmental and genetic components of LRS and assessing
110 evolvability in wild populations is of great contemporary importance. Between 30 to 50% of the
111 world's coral reefs have been lost and those remaining are considered highly vulnerable (Jackson
112 2010; De'ath *et al.* 2012; van Hooidonk *et al.* 2016). The rapid loss of suitable habitat is widely
113 acknowledged to be contributing to a decline in reef fish populations and biodiversity (Jones *et al.*
114 2004; Wilson *et al.* 2006; Paddock *et al.* 2009; Pratchett *et al.* 2018). Laboratory studies have
115 shown that near future environmental conditions predicted under climate change can have a
116 dramatic effect on reef fish reproductive success, and despite some levels of phenotypic plasticity
117 and transgenerational acclimation, the potential for adaptation is uncertain (Donelson *et al.* 2017;
118 Munday *et al.* 2013; Munday *et al.* 2017). To date, environmental, maternal and additive genetic
119 contributions to LRS in wild coral reef fish populations have not been assessed. However, recent

120 work establishing high levels of natal philopatry in some coral reef fishes (Jones *et al.* 2005,
121 D'Aloia *et al.* 2015; Salles *et al.* 2015; Almany *et al.* 2017) and the success of parentage analysis in
122 detecting family relationships across multiple generations (Salles *et al.* 2016a,b) opens the way for
123 quantifying [the local component of LRS](#) for the first time.

124
125 Here, we focus on the entire [local](#) population of the orange clownfish *Amphiprion percula* at Kimbe
126 Island, Papua New Guinea where each year ~half the juveniles successfully recruiting are progeny
127 of local breeding pairs (Salles *et al.* 2016a, Almany *et al.* 2017). We use multi-generational
128 pedigrees of up to 5 generations obtained from biennial DNA sampling over 10 years (Salles *et al.*
129 2016a,b) and apply a quantitative genetic linear mixed model approach (Kruuk & Hill 2008) to
130 quantify the additive genetic, maternal and environmental components of variation in LRS for the
131 self-recruiting portion of the population. Habitat effects were quantified by examining [this local](#)
132 LRS for individuals resident in two different anemone species and from different geographic
133 locations around the island (Salles *et al.* 2016a). By integrating habitat data with the pedigree
134 information in a quantitative genetic generalized linear mixed model, we were able to assess the
135 relative contribution of additive genetic, maternal and habitat effects to local LRS. We also
136 calculated the evolvability and heritability of [local](#) LRS to evaluate its evolutionary potential to
137 respond to selection at the scale [of the Kimbe island population](#).

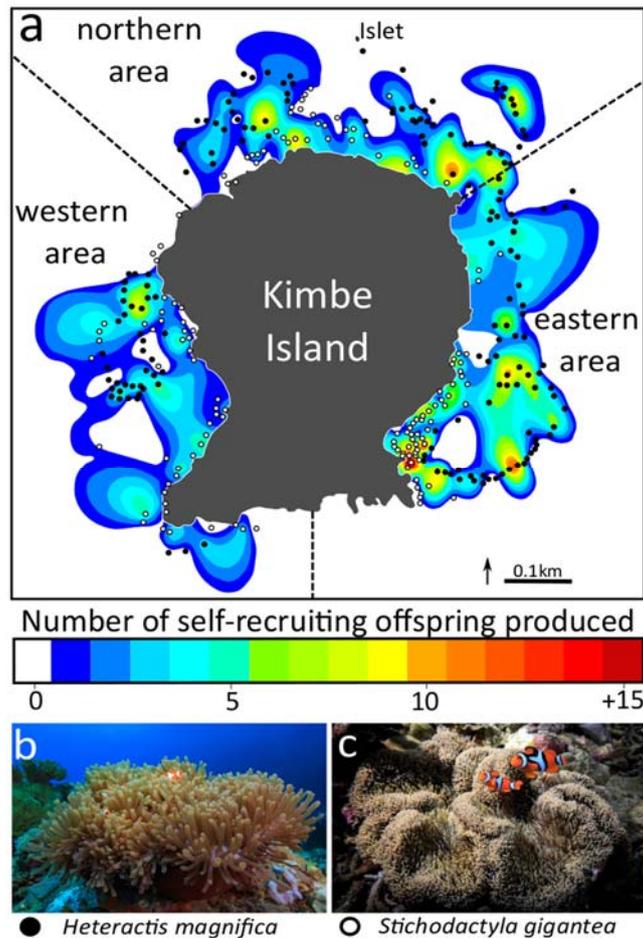
138

139 **METHODS**

140 **Study population and data collection**

141 A natural population of orange clownfish (*A. percula*) living in the reef surrounding Kimbe Island
142 (Fig. 1a; 5°12'22.56" S, 150°22'35.58" E), West New Britain Province of Papua New Guinea, was
143 surveyed every second year from 2003 to 2013. Here, *A. percula* lives in a mutualistic association
144 with one of two host sea anemone species, *Heteractis magnifica* (Fig. 1b) and *Stichodactyla*
145 *gigantea* (Fig. 1c). We geographically located and tagged a total of 310 anemones (176 *H.*

146 *magnifica* and 134 *S. gigantea*) that were occupied by *A. percula* on the entire reef surrounding the
 147 island.



148
 149 **Figure 1. Variation of the total number of offspring orange clownfish produced on each anemone around Kimbe**
 150 **Island between 2003 and 2013. (a)** The studied region was divided in three areas (northern, eastern and western areas).
 151 Colors correspond to the variation of the total number of juveniles locally self-recruited that were produced on each
 152 anemone (varying from 0 to 27) over a 10-year period. The expected value is interpolated from those around it (using
 153 default algorithms implemented in *Origin* software). Dots correspond respectively to the location of the two host
 154 anemones species: (b) *Heteractis magnifica* (black dots) and (c) *Stichodactyla gigantea* (white dots). Photos by Tane
 155 Sinclair-Taylor.
 156 These two anemone species are remarkably different in terms of the micro-habitat they provide,
 157 including a wide range of shapes, sizes, depth distributions and surrounding substrata (Dunn 1981;
 158 Chausson *et al.* 2018). Although we did not directly measure these variables (other than depth), the
 159 combination of host anemone species (*H. magnifica* or *S. gigantea*) with one of the three
 160 geographical areas covering the entire reef around the island (northern, western or eastern areas,

161 Figure 1a) where it is located describes a ‘habitat’ variable for each fish that encompasses a suite of
162 biotic and abiotic environmental conditions. These geographical areas correspond to the different
163 coasts of the island that reflect possible environmental effects of the geographic location (Salles *et*
164 *al.* 2016b). In total, the combination of the two host anemone species with the three geographical
165 areas allowed us to describe six different habitats.

166

167 Clownfish within one anemone live in group of typically three to five individuals in the Kimbe
168 Island population (Salles *et al.* 2015). The size-based dominance hierarchy in *Amphiprion* allows us
169 to determine the reproductive status of each individual (Fricke 1979). The female is the largest, the
170 male is the second largest, and the non-breeders rank progressively lower in the hierarchy as they
171 decrease in size. If the single female adult of a group dies, then the male changes sex to female, and
172 the largest non-breeder from the anemone becomes sexually mature as male. Reproduction occurs
173 year round, with females laying several hundred eggs in a clutch near the pedal disk of the host
174 anemone each lunar month. The eggs hatch after ~7 days of paternal care into larvae that spend ~10
175 days in the pelagic environment (Berumen *et al.* 2010) before settling [settling to an anemone, either](#)
176 [at their natal location \(Kimbe Island\) or elsewhere](#) (Planes *et al.* 2009).

177

178 Fish were captured by divers by using hand nets. Individuals were measured *in situ* using calipers,
179 fin-clipped (size > 35mm) or collected whole (size < 35mm) for genetic analysis and then released
180 back on the same anemone. Small pieces of fin tissue were preserved in 95% ethanol in 2-mL vials.
181 The biggest fish in each anemone was identified as the female, the second largest individual was
182 assumed to be the male, and all other individuals were classified as non-breeders. We extracted
183 DNA from all samples at 22 polymorphic microsatellite loci (Bonin *et al.* 2015). Then, we
184 identified the individuals sampled multiple times over the years by using the Excel macro GenAlex
185 v6.5 (Peakall & Smouse 2012) to compare multilocus genotypes from 2003, 2005, 2007, 2009,
186 2011 and 2013. Individuals were in average sampled 2.88 ± 0.04 times (mean \pm SE) over the six

187 surveys (1% of individuals persisted over the 10-yr period, Salles *et al.* 2016a). The 2-yr sampling
188 scheme precluded calculating a precise measurement of the age of individuals (Salles *et al.* 2016b),
189 in particular for fish sampled in 2003 during the first sampling period, which age was unknown.
190 The total duration of this long term survey did not allow us to obtain many replicated measurements
191 within individuals before and after sex change ($n = 41$ individuals). Estimating sex dependent
192 additive genetic variance is precluded in this case because some effects cannot be disentangled as
193 the clownfish only changes sex in one direction (from male to female). This change is always
194 associated with a change of sexual partner and with an increase in female body size, which we
195 expect to generate a confounding effect between a female condition and its genetic quality. We
196 therefore did not consider sex in our model as a result of data and analytical limitations.

197

198 **Pedigree used for quantitative genetic analysis**

199 Pedigree reconstruction was conducted by assigning juvenile fish to parental pairs on the basis of
200 their multilocus genotypes (Salles *et al.* 2016a). We used the software FaMoz (Gerber *et al.* 2003).
201 This approach is based on the calculation of LOD scores (Log of the odd-ratio comparison) for any
202 potential parentage relationship. It determined critical thresholds to accept or reject assignments by
203 simulating true and false parent-offspring pairs. Further details on parentage analyses and pedigree
204 reconstruction are given in Salles *et al.* (2016a). [We kept assignments to known parental pairs, but](#)
205 [rejected assignments to single adults.](#) In the context of overlapping generations, we used the year of
206 first sampling and the anemone of each parental couple as information to avoid possible false
207 assignments. As a result, sibship links could not be confused with parental links. Because the sex
208 changes through the life of the clownfish, the same individual can be related to its offspring with
209 either a paternal or maternal link. Based on the size of the two parents and the year of first capture,
210 we can identify the mother and the father. The original population pedigree includes 2927 clownfish
211 over five generations including 121 families, 987 paternal, 987 maternal, 1809 full-sib, 412
212 maternal half-sibs, 248 paternal half-sib, 135 maternal grandmothers, 135 maternal grandfathers,

213 278 paternal grandmothers, 278 paternal grandfathers and 218 cousins (Salles *et al.* 2016a). For this
214 study, we excluded from the original pedigree the 1192 individuals that were removed from the
215 habitat at the juvenile stage (size < 35mm, 10 to 458 days old). The final pedigree used for this
216 study includes 1735 individuals from five generations (Fig. S1). We used the R package ‘pedantics’
217 (Morrissey & Wilson 2010) to assess the power of the resolved pedigree to detect significant
218 quantitative genetic parameters (Fig. S1).

219

220 **LRS: the individual contribution to self-recruitment**

221 **LRS, which when measured at the scale of the local population is also the** contribution of an
222 individual to self-recruitment, corresponds to the total number of offspring produced during its
223 lifetime and recruiting into Kimbe Island (*e.g.*, the local breeder population). To deal with the fact
224 that some fish were still alive at the end of sampling and that some fish might have already
225 reproduced before the first sampling year, we used biennial measurements of their reproductive
226 success (using field-data from 2003, 2005, 2007, 2009, 2011 and 2013) to compare LRS between
227 individuals. The LRS corresponds here to the total number of descendants produced on a biennial
228 basis that successfully recruited into Kimbe Island population, which provided us with repeated
229 measures over the period of the survey from 2003 to 2013. In the Supplementary information we
230 present results from an alternative approach based on the De-lifing method (DL). The calculation of
231 DL takes into account the temporal variation of the population growth and estimates the
232 contribution of every clownfish to biennial changes in population size through both reproduction
233 and survival (Coulson *et al.* 2016). Statistical problems potentially leading to precision issues and
234 invalid conclusions have been associated with the use of DL (Dupont *et al.* 2017) but DL has only
235 been used in two of the 15 studies where the genetic variation of fitness was quantified in wild
236 populations (Table S1), which limits our ability to discuss its properties. **We therefore also provided**
237 **DL results in this study in the supplementary section.**

238

239 **Quantitative genetic generalized linear mixed model approach**

240 Similarities between relatives living in contrasted micro-habitats allowed us to evaluate
 241 simultaneously the genetic and habitat components of LRS. Repeated ‘records’ on individuals made
 242 it possible to estimate permanent environmental effects, which allowed us to account for intra-
 243 individual and unmeasured environmental trait variation across time. Permanent environmental
 244 effects also account for a part of non-additive genetic effects (Wilson *et al.* 2010). The LRS
 245 variance was partitioned into six random effects: Additive genetic (V_A), Maternal (V_M), Natal
 246 Habitat (V_{NH}), Resident Habitat (V_{RH}), Permanent Environment (V_{PE}) and Residual (V_R) variances
 247 by using the ‘animal model’ quantitative genetic approach (Kruuk 2004). This Linear Mixed Model
 248 (LMM) approach uses pedigree information to extract the additive genetic component. This
 249 approach is more powerful than traditional analyses (*e.g.*, parent–offspring regressions) because it
 250 takes into account every relationship link in a pedigree. Maternal variance was modeled using the
 251 mother’s identity as a random effect, allowing maternal effects to include both genetic and
 252 environmental maternal effects. Permanent environmental effects were modeled by including the
 253 identity of individuals as a random effect. The LRS variance is the sum of six variance components:

$$254 \quad V_{LRS} = V_A + V_M + V_{NH} + V_{RH} + V_{PE} + V_R \quad (1)$$

255 Quantitative genetic models were computed as univariate GLMMs using the ‘MCMCglmm’
 256 package (Hadfield 2010) in R version 3.5.1 (R.Core.Team 2018), with LRS as a Poisson response
 257 variable. Using this Bayesian framework facilitated parameter estimation for non-Gaussian traits.
 258 We used parameter expanded priors for all analyses ($V=1$, $nu=0.02$), which are often referred to as
 259 ‘non informative’ priors although such denomination can be debated, as we wanted posterior
 260 estimates to be determined from the data and not from the priors (Morrissey *et al.* 2014). We ran
 261 model MCMC chains over 1,000,000 iterations with initial burning of 10,000 iterations and a
 262 thinning of 1,000 iterations. Historically, the Deviance Information Criterion (DIC) was often used
 263 to compare models and assess the significance of the random variance components in this type of
 264 approach. However, it is becoming less commonly used since it was recognized as an inappropriate

265 tool for model comparisons of the same type than quantitative genetic GLMM analyses (Gelman *et*
 266 *al.* 2014; Spiegelhalter *et al.* 2014). Effects of variance components were considered statistically
 267 supported if their posterior distributions did not overlap zero (Wilson *et al.* 2010).

268

269 **Variance Components**

270 Variance components were estimated as the mode of the posterior distributions established on the
 271 MCMC sample and we reported the lower and the upper limits of the 95% credible interval. For the
 272 six variance components, we calculated their relative contribution to the sum of all variance
 273 components, thereby expressing their effects as percentages of the total phenotypic variance (V_{LRS}).
 274 As a result, we obtained standard narrow sense heritability estimates for LRS (h^2) by applying the
 275 basic formula ($h^2 = V_A / V_P$, see Falconer & McKay 1996), and similarly maternal effects by
 276 estimating the proportion of total phenotypic variance explained by the maternal variance ($m^2 = V_M /$
 277 V_P). Evolvability (I_A) of LRS, equal to the additive genetic variance divided by the squared mean of
 278 the LRS (Wagner & Altenberg 1996), was estimated to evaluate the capacity for adaptive
 279 evolutionary change of the number of offspring that self-recruit in the population. The analyses
 280 conducted in the ‘MCMCglmm’ framework assumed a Poisson distribution and therefore provided
 281 parameter estimates for evolutionary inference or future comparisons on a statistically convenient
 282 latent scale for non-Gaussian traits. We therefore endeavored to back-transform all the estimates of
 283 the latent scale variables included in the model (see equation 1) onto the observed data scale to
 284 improve our inferences. We used the ‘QGlmm’ package (de Villemereuil *et al.* 2016) to back-
 285 transform the estimates, specifically the function ‘QGparams’ to estimate additive components such
 286 as V_A and h^2 , and ‘QGicc’ to estimate broader sense components such as V_M and m^2 , V_{NH} , V_{RH} , V_{PE}
 287 and V_R . Although parameter estimates transformed back on the data-scale are expected to be upward
 288 biased, their ratio is reliable, and hence the estimators derived from their relative proportions such
 289 as h^2 . It is necessary to point out two specific aspects of the back transformation on the observed
 290 data scale. First, V_R is estimated on the basis of the additive over-dispersion term in the nonlinear

291 model and its value cannot be interpreted similarly to the usual residual variance term estimated by
292 classical quantitative genetics generalized linear mixed models. Second, the sum of the variance
293 components estimated on the data scale are not additive and therefore not expected to sum up to the
294 value of the phenotypic variance calculated directly on the raw data. For the sake of clarity and
295 comparison, we present the results on the latent scale and the observed data scale. We calculated the
296 95% credibility intervals from the posterior distributions of observed parameters for all the variance
297 components and other estimates expressed on their basis by using the ‘HDInterval’ package
298 (Meredith & Kruschke 2016).

299

300 RESULTS

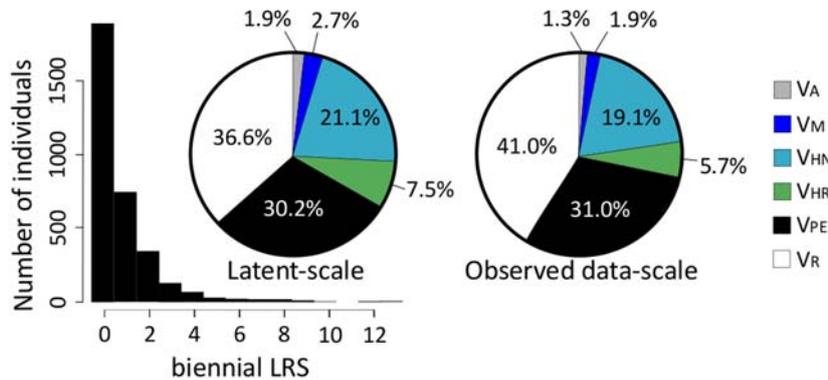
301 Habitats dominantly shape the Lifetime Reproductive Success in the clownfish population

302 Biennial estimates of the Lifetime Reproductive Success (LRS) [estimated inside the area of the](#)
303 [Kimbe island population](#) ranged from 0 to 13, with a phenotypic variance $V_{LRS}=1.31$ and an average
304 value of 0.54 ± 0.05 (mean \pm SE) offspring per individual for a two-year period. Because clownfish
305 live in strong association with their anemone, we were able to identify and geo-locate the precise
306 position and habitat where breeders contributed more to the local replenishment of the population
307 (Fig. 1a). For example, breeders that produced more self-recruiting offspring lived in Kimbe
308 Island’s eastern area and mostly in *S. gigantea* anemones. Our analysis also revealed fish that did
309 not contribute at all to the local replenishment of the population over the 10-year monitoring period.
310 These fish were located in 48 *H. magnifica* and 30 *S. gigantea* of the 310 anemones monitored in
311 both deep and shallow waters (Fig. 1a). We found that 25% of the pairs of local breeders did not
312 contribute at all to the renewal of the population over a period of 10 years.

313

314 Quantitative genetic linear mixed models on the latent and the observed data scale gave very close
315 results. Our results on the scale of observed data showed that Natal Habitat and Resident Habitat
316 explained respectively 19.1% and 5.7% of the variance in [local](#) LRS, furthermore, residual and

317 permanent environment explained respectively 41.0% and 31.0%, whereas additive genetic effects
 318 and maternal effects were very weak and explained 1.3% and 1.9% respectively (Fig. 2a, variances
 319 on observed data-scale). Similar results were obtained for DL (see Supplementary Information for
 320 more details).



321

322 **Figure 2. Sources of variation in the biennial estimate of the local Lifetime Reproductive Success (LRS) of the**

323 **Kimbe Island orange clownfish.** Distribution of the biennial estimate of the LRS (histograms). Variance components

324 on both latent-scale and observed data-scale (pie charts) for the biennial estimate of the LRS explained by Additive

325 genetic (V_A), Maternal (V_M), Natal Habitat (V_{NH}), Resident Habitat (V_{RH}), Permanent Environmental (V_{PE}) and

326 Residual (V_R) variances. These proportions were calculated from the values of the posterior modes of a quantitative

327 genetics generalized linear mixed model analysis (for details see Table 1).

328

329 **Low evolvability and low heritability for LRS**

330 The modes of the posterior distributions estimating additive genetic variance, expressed on

331 observed data-scale, were extremely small for the two measures of self-recruitment (Table 1).

332

333 **Table 1. Sources of variation in Lifetime Reproductive Success (LRS) for the Kimbe Island orange clownfish.**

334 Here we reported variance component estimates quantified by using the animal model approach: Additive genetic (V_A),

335 Maternal (V_M), Natal Habitat (V_{NH}), Resident Habitat (V_{RH}), Permanent Environmental (V_{PE}) and Residual (V_R)

336 Variances. We also report size effects as proportions of explained phenotypic variance: narrow-sense heritability (h^2),

337 maternal effects (m^2) and the mean standardized additive genetic variance: evolvability (I_A) for biennial LRS. Measures

338 are expressed on a latent-scale (direct *MCMCglmm* R results) and the observed data-scale (*QGglmm* R back-

339 transformation). 95% credible intervals (CI) are reported for each estimate.

	LRS	LRS
	Latent scale	Observed data-scale
V_A	0.046	0.030
(CI)	(1.381×10^{-3} to 0.146)	(4.94×10^{-4} to 0.060)
V_M	0.067	0.046
(CI)	(2.000×10^{-3} to 0.211)	(8.822×10^{-3} to 0.287)
V_{NH}	0.516	0.450
(CI)	(0.015 to 1.529)	(0.126 to 1.524)
V_{RH}	0.184	0.135
(CI)	(0.264 to 0.473)	(0.038 to 0.457)
V_{PE}	0.737	0.726
(CI)	(0.496 to 0.952)	(0.203 to 2.460)
V_R	0.894	0.963
(CI)	(0.717 to 1.105)	(0.270 to 3.264)
V_{LRS}	2.44	2.35
(CI)	(1.71 to 3.65)	(0.65 to 8.05)
h^2	0.019	0.013
(CI)	(6.827×10^{-4} to 0.057)	(4.951×10^{-5} to 1.227×10^{-2})
m^2	0.027	0.019
(CI)	(9.157×10^{-4} to 0.083)	(2.966×10^{-5} to 2.044×10^{-2})
I_A	0.154	0.103
(CI)	(4.643×10^{-4} to 0.492)	(1.661×10^{-3} to 0.511)

340

341 We found $V_A=0.030$ (CI_{95%} 4.94×10^{-4} to 0.060). This could be linked to the statistical power of our
342 pedigree (Fig. S1). Our model nevertheless placed fairly restricted bands on the 95% credible
343 intervals (Table 1). Credible intervals did not overlap the zero but were close. The extent to which
344 these very low values of genetic estimates are not null must therefore be considered with caution.
345 LRS evolvability estimated on the observed data-scale, which evaluated the micro-evolutionary
346 change of the number of self-recruiting offspring that can be reached by the population, was equal
347 to 0.103 (CI_{95%} 1.661×10^{-3} to 0.511). In other words, 0.104 additional juveniles were added to the

348 average number of juveniles originating and recruiting in the population per generation. The
349 heritability estimate expressed on the observed data-scale was $h^2=0.013$ (CI_{95%} 4.951×10^{-5} to
350 1.227×10^{-2}) for LRS (Table 1). We can therefore estimate the maximum response (R) to selection
351 (S), in the presence of strong selection pressures acting on the Kimbe Island orange clownfish
352 population by using the Breeder's equation $R=h^2 \times S$ (Falconer & Mackay 1996; Lush 2008). The
353 low to negligible value of the LRS heritability means that the maximum genetic change of the
354 population average LRS would never exceed ~ 0.020 offspring per generation. Similar results were
355 obtained for DL (see Supplementary Information for more details).

356

357 **Low maternal effects for LRS**

358 We found that maternal variance for the LRS was extremely small, to the extent that it might be
359 considered as null (Table 1): $V_M=0.019$ (CI_{95%} 2.966×10^{-5} to 2.044×10^{-2}). While our analysis
360 detected maternal variance, it made very little contribution to the total variance in LRS ($m^2=1.9\%$,
361 expressed on observed data-scale, Fig. 2). The habitat occupied by the mother (Natal Habitat) had a
362 stronger effect on LRS than the mother herself. The relative contribution of individuals to the
363 population replenishment was indeed influenced by the Natal Habitat to an extent of 19.1% for
364 LRS. Similar results were obtained for DL (see Supplementary Information for more details).

365

366 **DISCUSSION**

367 *Strong habitat and weak genetic effects on LRS*

368 Our study revealed that LRS in the Kimbe Island orange clownfish population quantified over five
369 generations was largely explained by host anemone species and geographical location ($\sim 97\%$), with
370 only weak maternal (1.9%) and additive genetic effects (1.3%). The strong effects of habitat can be
371 attributed to the intrinsic biological characteristics of the anemone species (*e.g.*, size, shape and
372 toxicity) and their effects of on life-history traits of their resident clownfish (Salles *et al.* 2016b;
373 Chausson *et al.* 2018). In addition, the higher toxicity of *S. gigantea* (Nedosyko *et al.* 2014) might

374 provide better protection against predators of eggs attached near the pedal disk of the host anemone,
375 but this hypothesis remains to be tested. The geographical location of the different host-anemones
376 also appears to be important, with more successful individuals in shallow water, close to the land on
377 *S. gigantea* and in deeper lagoons for *H. magnifica*, which might promote greater local retention of
378 larvae. To date, the mechanisms responsible for geographical differences in LRS around Kimbe
379 Island remain unknown (Berumen *et al.* 2010). It is likely that some breeders have a different
380 reproductive success beyond the sampled population, through dispersers rather than self-recruiters.
381 The weak genetic effects on LRS that shows a very low to negligible rate of adaptation inside the
382 Kimbe island population raise concern about the ability of this reef fish population to exhibit
383 longer-term adaptive evolution in response to rapid climate change.

384

385 *Susceptibility to habitat change*

386 Habitat is clearly the main driver of the variation in LRS and individuals that happen to settle on
387 particular anemones and particular places do well. The dependence of LRS on habitat quality inside
388 the Kimbe island population suggests this species will be extremely susceptible to habitat
389 degradation over ecological time scales. Direct and indirect human impacts on clownfish
390 populations and their anemone habitats are already affecting the habitat quality for numerous
391 clownfish species (Jones *et al.* 2008; Saenz-Agudelo *et al.* 2011; Bonin *et al.* 2016). *S. gigantea*
392 anemones located in shallow waters are likely to be disproportionately more impacted by increasing
393 water temperatures and irradiance (Bak *et al.* 2005; Hobbs *et al.* 2013). If these locations are
394 differentially impacted, this will affect the individual contribution to the local replenishment of the
395 population and compromise population persistence.

396

397 *Low to negligible evolutionary potential*

398 Our findings provide the first empirical support for a wild marine population to Fisher's
399 fundamental theorem of selection that additive genetic variance in fitness is depleted under

400 selection and tends towards zero in a population reaching evolutionary equilibrium (Fisher 1930;
401 Gustafsson 1986; Falconer 1989). Although normal and expected, low heritability and evolvability
402 in LRS is concerning given the increasing rate of environmental change. The low to negligible
403 scope for adaptive evolution (estimated by evolvability) and the low to negligible genetic potential
404 for responding to selection (estimated by heritability) may not be a problem for gradual
405 environmental change. At this rate, it would take around at least 10 generations for the population
406 average LRS to increase by one juvenile, which highlights the stability of the demographic rate of
407 self-recruitment in this population. Our results therefore support the hypothesis that the population
408 is at evolutionary equilibrium (no genetic changes) in a context of environmental stability over the
409 timescale of the survey.

410

411 *Connectivity as a plausible cause*

412 Our finding that the Kimbe Island clownfish population harbored low to negligible additive genetic
413 variation for LRS over a ten-year period was at first surprising because immigration accounts for on
414 average 44% of the juvenile recruitment (Salles *et al.* 2015; 2016a). The average dispersal distance
415 in Kimbe Bay is between 10 and 20km, providing substantial connectivity among adjacent reefs and
416 potential dispersal of up to 100km (Almany *et al.* 2017; Pinsky *et al.* 2017). The associated gene
417 flow would be expected to bring new genetic variants and thereby increase genetic variation for
418 LRS (Keller *et al.* 2001; Lavergne & Molofsky 2007; Facon *et al.* 2008). Under such scenario,
419 selection for self-recruitment, and thereby against migrants, would have to be strong to keep the
420 population at evolutionary equilibrium. An alternative scenario is that homogenization by gene flow
421 results in most immigrants sharing a similar genetic background. As a result, low genetic variation
422 would be maintained because no new genetic variants were brought in the population by gene flow
423 (Pujol *et al.* 2010). Low genetic variation for LRS implies that evolution by selection at the local
424 scale is extremely limited in its current state. However, this does not imply a dead end for the

425 adaptive evolution of this population because several mechanisms can provide adaptive
426 evolutionary potential over the long term (Pujol *et al.* 2018).

427

428 *Slight but probably negligible maternal genetic effect*

429 Additional adaptive evolutionary potential can in theory be provided to a population by maternal
430 effects (Räsänen & Kruuk 2007). In the Kimbe Island orange clownfish, we found that maternal
431 effects explained up to 2.7% of the LRS variance, which is quite small, even if it was more than
432 additive genetic effects. Our findings thereby revealed that maternal effects increased slightly the
433 low to negligible rate of LRS change by adaptive evolution. One should note that this increase was
434 nearly negligible. It is likely that this low value represents the genetic component of the maternal
435 effect because the identity of each mother was taken into account. The habitat of birth, on the other
436 hand, which is also the maternal habitat, might encompass some direct effect of the maternal
437 environment (Germain & Gilbert 2014). The environmental component can represent a non-
438 negligible part of parental effects (Chirgwin *et al.* 2017). In fact, there is growing awareness that
439 maternal environmental effects can contribute to adaptation in natural populations, especially when
440 maternal and offspring environments are positively correlated (Burgess & Marshall 2014; Shama
441 2015; Dey *et al.* 2016). It might even buy some time for adaptive evolution through slow genetic
442 change to occur (Levis & Pfenning 2016).

443

444 *Towards a wider sample of contemporary rates of adaptive evolution in the wild*

445 In our study, LRS estimates the individual fish contribution to the local population replenishment.
446 This excludes the dispersal fitness because the amount of offspring produced and which dispersed
447 to live somewhere else is unknown on wild population pedigrees (Kruuk *et al.* 2000; Merilä &
448 Sheldon 2000, McCleery *et al.* 2004). Its genetic variation evaluates the rate of adaptive evolution
449 [inside the Kimbe island population](#). While there are no comparable data from marine systems, 15
450 studies conducted on terrestrial vertebrates have also estimated the additive genetic variation and

451 the heritability of LRS (Table S1). It is noteworthy that the number of estimates of maternal effects
452 on LRS variation are extremely rare (Kruuk *et al.* 2000, Foerster *et al.* 2007; Schroeder *et al.* 2012;
453 McFarlane *et al.* 2014). A majority of these studies similarly found low to negligible contributions
454 of additive genetic effects, while the situation is less clear for maternal effects, partly because
455 studies remain scarce. The existence of additive and maternal genetic variation for fitness, even
456 when very low, implies that the population was not totally at equilibrium because there was a small
457 genetic change in the wild population over the course of the long term survey. It also implies very
458 limited genetic adaptive potential.

459

460 **CONCLUSION**

461 The major outcome of this study is that the heterogeneity of the habitat of the Kimbe Island orange
462 clownfish had a profound influence on the individual contribution to the local population
463 replenishment over five generations. This finding implies that habitat ecology is crucial for this
464 clownfish population. In terms of future persistence, expected changes in habitat quality and
465 configuration over relatively short time scales might affect the ability of fish to self-recruit. This
466 ability harbored low to negligible additive genetic and maternal genetic variation. As a
467 consequence, this population potential for rapid evolutionary change of LRS by selection, and
468 therefore its rate of adaptive evolution, can be considered negligible in the current state of the
469 population. This finding, which is in line with other studies on the topic, stresses the importance of
470 environmental mechanisms (*e.g.*, plasticity) that have the potential to enable rapid adaptive
471 responses (Donelson *et al.* 2017; Munday *et al.* 2017). Our findings suggest a further evaluation of
472 maternal environmental effects is needed to better evaluate their role in the resilience of wild
473 populations (Shama 2015; Chirgwin *et al.* 2017). From the perspective of management, our results
474 caution against hoping for local adaptive responses and lend support to focusing conservation
475 efforts on maintaining habitat quality.

476

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491 **AUTHORSHIP**

492 GPJ, SLRT, and SP designed the research program; OCS, BP, GRA, and MLB contributed new
493 reagents/analytic tools; OCS and BP analyzed data; OCS, BP et SP wrote the manuscript and all
494 authors contributed substantially to revise the paper.

495 **DATA ACCESSIBILITY STATEMENT**

496 Data from this study can be obtained by using the CRIOBE data portal (<http://www.criobe.pf>). R
497 programming protocol can be obtained on the Zenodo repository **DOI**

498 **CONFLICT OF INTEREST DISCLOSURE**

499 The authors of this preprint declare that they have no financial conflict of interest with the content
500 of this article.

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748 population subject to natural immigration and inbreeding. *Evolution* 72-10, 2057-2075.

749 **SUPPORTING INFORMATION**

750 Additional Supporting Information may be found in the online version of this article.

751 The Additional Supporting Information file contains:

- 752 • **Table S1.** Previous estimates of fitness heritability and maternal effects on fitness in 15 wild
753 populations.
- 754 • **Figure S1.** Pedigree data and power analysis for the Kimbe Island orange clownfish
755 population.
- 756 • **Supplementary methods, results and discussion** on De-lifing measures.

757

1 **Supporting Information**

2 **Strong habitat and weak genetic effects shape the lifetime reproductive success**
3 **in a wild clownfish population**

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5 Agudelo, Maya Srinivasan, Simon R. Thorrold, Benoit Pujol*, and Serge Planes*

6
7 **Both authors share senior authorship of this article*

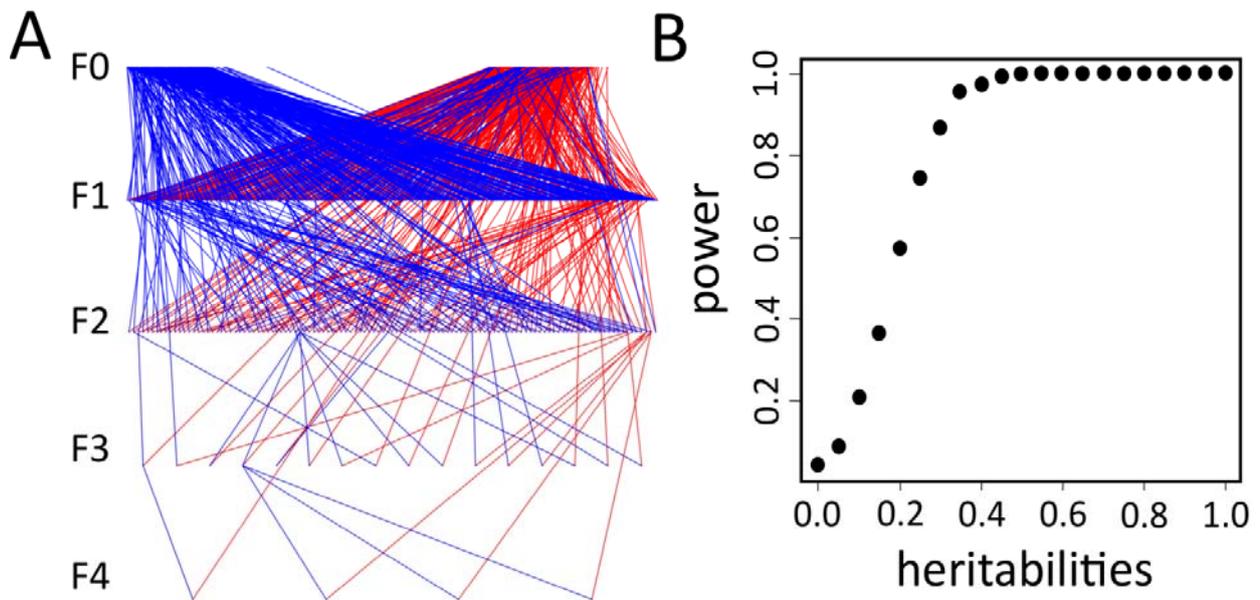
8 **Correspondence:** Océane Salles and Benoit Pujol; Emails: oceane.salles@gmail.com and
9 benoit.pujol@univ-perp.fr

10
11 **This file contains:**

- 12 • **Table S1.** Previous estimates of fitness heritability and maternal effects on fitness in 15 wild
13 populations.
- 14 • **Figure S1.** Pedigree data and power analysis for the Kimbe Island orange clownfish
15 population.
- 16 • **Supplementary methods, results and discussion** on De-lifing measures.

17 **Table S1. Previous estimates of fitness heritability and maternal effects on fitness in wild populations.** Different fitness measures are denoted as
 18 lifetime reproductive success (LRS), De-lifing or relative RS (LRS/mean of LRS in the population). ♀ corresponds to female, ♂ corresponds to male
 19 and ♀♂ corresponds to female and male confounded. NA indicates that maternal effects were not estimated or dropped from the quantitative genetic
 20 models.

Study organism	Location	Years of monitoring	Heritability estimate ± SE or 95%SI	Maternal effects estimate	Fitness measure	Reference
Collared flycatcher	Gotland Is.	1980 to 1984	♀ 0.01 ± 0.16 ♂ 0.01 ± 0.13	NA NA	LRS	(Gustafsson 1986)
Red deer	Isle of Rum	1971 to 1999	♀ 0.00 ± 0.05 ♂ 0.02 ± 0.06	0.16 ± 0.041 NA	LRS	(Kruuk <i>et al.</i> 2000)
		1971 to 2005	♀ 0.21 ± 0.06 ♂ 0.07 ± 0.06	0.0021 0.0045	De-lifing	(Foerster <i>et al.</i> 2007)
Collared flycatcher	Gotland Is.	1980 to 1997	♀ 0.21 ± 0.06 ♂ 0.07 ± 0.06	NA NA	LRS	(Merila & Sheldon 2000)
Bighorn sheep	Ram Mountain	1973 to 1998	♀ 0.66 ± 0.32	NA	LRS	(Réale & Festa-Bianchet 2000)
	Sheep River	1981 to 1998	♀ 0.19 ± (0.50)	NA	LRS	
Cheetahs	Serengeti park	1970 to 1994	♀ 0.88 ± ?	NA	LRS	(Kelly 2001)
Great tit	Wytham Wood	1960 to 1998	♀ 0.00 ± 0.04 ♂ 0.02 ± 0.04	NA NA	LRS	(McCleery <i>et al.</i> 2004)
Bighorn sheep	Ram Mountain	1971 to 2003	♀ 0.00 ± 0.00 ♂ 0.00 ± 0.00	NA NA	LRS	(Coltman <i>et al.</i> 2005)
Red billed gulls	Kaikoura	1958 to 2004	♀ 0.36 ± 0.29 ♂ 0.00 ± 0.00	NA NA	LRS	(Teplitsky <i>et al.</i> 2009)
Rhesus macaques	Cayo Is.	1959 to 1990	♀ 0.36 ± 0.081	NA	LRS	(Blomquist 2010)
House sparrow	Lundy Is.	2000 to 2011	♀ 0.09 (0.03 to 0.18)	0.33 (0.14 to 0.51)	De-lifing	(Schroeder <i>et al.</i> 2012)
Soay sheep	St Kilda	1985 to 2002	♀ 0.03 ± 0.01	NA	Relative RS	(Morrissey <i>et al.</i> 2012)
Red squirrels	Yukon	1987 to 2011	♀ 4.90x10 ⁻⁰⁴ (3.0x10 ⁻⁰⁸ to 0.07) ♂ 6.80 x10 ⁻⁰⁴ (8.5x10 ⁻¹¹ to 0.10) ♀♂ 4.90 x10 ⁻⁰⁴ (1.1x10 ⁻⁰³ to 0.39)	0.07 (0.02 to 0.14) 0.08 (0.01 to 0.14) 0.10 (0.10 to 0.37)	LRS	(McFarlane <i>et al.</i> 2014)
Savannah sparrow	Kent Is.	1987 to 2005	♀ 0.002 ± 0.036 ♂ 0.000 ± 0.036	NA NA	LRS LRS	(Wheelwright <i>et al.</i> 2014)
Brown anole lizard	Kidd Cay	2005 to 2008	♀ 1.40 (8.0x10 ⁻¹⁰ to 0.023)	NA	RS	(Calsbeek <i>et al.</i> 2015)
Song sparrow	Mandarte Island	1993-2015	♀ 0.5 ± 0.21 ♂ 0.44 ± 0.74	NA NA	LRS	(Wolak <i>et al.</i> 2018)



22

23 **Figure S1.** Pedigree data of the Kimbe Island orange clownfish population. **(A)** Pedigree
 24 representation of the orange clownfish *Amphiprion percula* in Kimbe Island (n= 1735, excluding
 25 new-recruits). Each line connects a parent with one of its offspring (blue and red lines represent
 26 respectively paternal and maternal links). It is important to note that an individual can be father then
 27 mother. The generation is indicated on the left from first generation (F0, n=502) to fifth generation
 28 (F4). **(B)** Power analysis of the pedigree.

29 **SUPPLEMENTARY METHODS**

30 **Individual's contribution to biennial population growth rate**

31 We estimated the De-lifing (DL), which is the individual's contribution to the biennial population
 32 growth rate. DL was calculated following Coulson *et al.* (2006). This index can be considered as the
 33 realized fitness of an individual over the given period of time (here two years) by including survival
 34 and success to local recruitment. This index offers great opportunities to empirically study
 35 ecological and evolutionary changes in stochastic environments (Dupont *et al.* 2017).

36 According to Coulson *et al.* (2006), an individual's biennial fitness is measured by its contribution
 37 DL_{ti} to the growth rate of the population between t and $t + 1$.

38
$$DL_{ti} = \frac{s_{ti} - \bar{s}_t}{N_{t-1}} + \frac{F_{ti} - \bar{F}_t}{N_{t-1}} \quad (1)$$

39 where N_t is the population size at time t ; s_{ti} is the survival of individual i between t and $t + 1$ (1 if
 40 it survived and 0 otherwise); \bar{s}_t is the mean survival rate in the population between t and $t + 1$; F_{ti}
 41 is the fecundity of the individual i defined as the number of offspring born between t and $t + 1$ and

42 still alive at $t + 1$ and \bar{F}_t is the mean individual fecundity in the population. A negative value of DL
43 represents an individual that performed worse than the population mean while a positive value
44 represents one that performed better.

45 **Quantitative genetic generalized linear model approach**

46 As for the IIRS variance, the DL variance was partitioned into six random effects: Additive genetic
47 (V_A), Maternal (V_M), Natal Habitat (V_{NH}), Resident Habitat (V_{RH}), Permanent Environment (V_{PE})
48 and Residual (V_R) variances by using the ‘animal model’ quantitative genetic approach (Kruuk
49 2004).

50 The DL variance is the sum of six variance components:

$$51 \quad V_{DL} = V_{A_{DL}} + V_{M_{DL}} + V_{NH_{DL}} + V_{RH_{DL}} + V_{PE_{DL}} + V_{R_{DL}} \quad (2)$$

52 We used a quantitative genetics model as univariate GLMMs using the 'MCMCglmm' package
53 (Hadfield 2010) in R version 3.5.1 (R Core Team 2018), with DL as a Gaussian response variable.

54 We used parameter expanded priors for all analyses ($V=1, nu=0.02$), which are often referred to as
55 ‘non informative’ priors although such denomination can be debated, as we wanted posterior
56 estimates to be determined from the data and not from the priors (Morrissey *et al.* 2014). We ran
57 model MCMC chains over 1,000,000 iterations with initial burning of 10,000 iterations and a
58 thinning of 1,000 iterations.

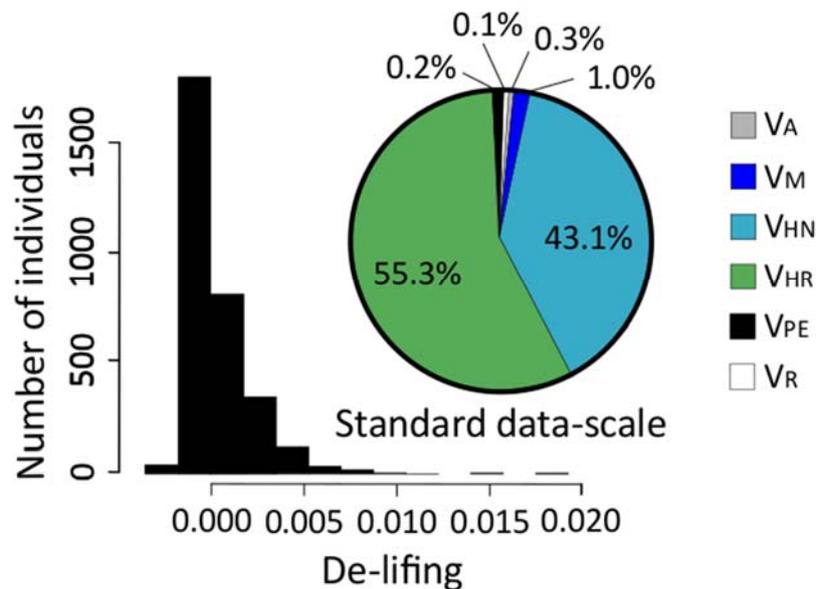
59 Variance components were estimated as the mean of the posterior distributions (the mean of the
60 MCMC sample). and we reported the lower and the upper limits of the 95% credible interval. For
61 the six variance components, we calculated their relative contribution to the sum of all variance
62 components, thereby expressing their effects as percentages of the total phenotypic variance (V_{DL}).
63 As a result, we obtained standard heritability estimates for DL (h^2_{DL}) by applying the basic formula
64 ($h^2 = V_A / V_P$, see Falconer & McKay 1996), and similarly maternal effects by estimating the
65 proportion of total phenotypic variance explained by the maternal variance ($m^2 = V_M / V_P$). Because
66 DL is Gaussian variable, we reported a standard heritability and a standard maternal effects. We
67 calculated the 95% credibility intervals from the posterior distributions of observed parameters for

68 all the variance components and other estimates expressed on their basis by using the ‘HDInterval’
69 package (Meredith & Kruschke 2016).

70 SUPPLEMENTARY RESULTS

71 Habitats dominantly shape the De-lifing in the clownfish population.

72 De-lifing (DL) varied from -2.56×10^{-3} to 1.81×10^{-2} (Figure S2, histogram), which illustrates that
73 there is variation in the individual contribution to the population demographic growth through
74 reproduction and survival. Its average value at the population level was $2.16 \times 10^{-5} \pm 8.04 \times 10^{-5}$ units.
75 When considering DL, 43.1% of its variance was explained by Natal Habitat (Figure S2, pie chart)
76 and 55.3% by Resident Habitat, again with minor contributions from Additive genetic effects
77 (0.3%) and Maternal effects (1.0%).



78

79 **Figure S2.** Sources of variation in De-lifing (DL) variance of the the Kimbe Island orange
80 clownfish. Distribution (histograms) and variance components on standard data-scale (pie chart) for
81 biennial DL explained by Additive genetic (V_A), Maternal (V_M), Natal Habitat (V_{NH}), Resident
82 Habitat (V_{RH}), Permanent Environment (V_{PE}) and Residual (V_R) variances. These proportions were
83 calculated from the values of the posterior mode of a quantitative genetics generalized linear mixed
84 model analysis.

85 We found that Additive genetic variance was extremely low: $V_{A,DL} = 3.151 \times 10^{-5}$ ($CI_{95\%}$ 2.897×10^{-5}
86 to 3.470×10^{-5} , Table S2). Consequently, heritability estimates expressed on the standard data-scale
87 was $h^2_{DL} = 0.003$ ($CI_{95\%}$ 2.432×10^{-3} to 0.0261, Table S2) for DL. We also found that maternal was

88 extremely small, to the extent that it might be considered as null: $V_{M,DL}=1.190\times 10^{-4}$ (CI_{95%}
 89 9.974×10^{-5} to 1.434×10^{-4} , Table S2). While our analysis detected maternal variance, it made very
 90 little contribution to the total variance in DL estimate ($m^2_{DL}=1.0\%$, Figure S2). The relative
 91 contribution of individuals to the population replenishment was indeed influenced by the Natal
 92 Habitat to an extent of 43.1% for DL.

93 **Table S2. Sources of variation in De-lifing (DL) for the Kimbe Island orange clownfish.** Here
 94 we reported variance component estimates quantified by using the animal model approach: Additive
 95 genetic effect (V_A), Maternal effect (V_M), Natal Habitat (V_{NH}) and Resident Habitat (V_{RH}) effects,
 96 Permanent Environmental (V_{PE}) effect and Residual Variance (V_R) and heritability (h^2) and maternal
 97 effects (m^2) for biennial DL. 95% credible intervals (CI) are reported for each estimate.

Components	DL Standard scale
V_A (CI)	3.151×10^{-5} (2.897×10^{-5} to 3.470×10^{-5})
V_M (CI)	1.190×10^{-4} (9.974×10^{-5} to 1.434×10^{-4})
V_{NH} (CI)	5.022×10^{-3} (9.972×10^{-4} to 0.012)
V_{RH} (CI)	6.443×10^{-3} (1.13×10^{-3} to 0.017)
V_{PE} (CI)	3.016×10^{-5} (2.72×10^{-5} to 3.25×10^{-5})
V_R (CI)	2.978×10^{-5} (1.15×10^{-5} to 1.29×10^{-5})
h^2 (CI)	0.003 (6.506×10^{-4} to 6.712×10^{-3})
m^2 (CI)	0.010 (2.432×10^{-3} to 0.0261)

98

99 SUPPLEMENTARY DISCUSSION

100 Because of the different proprieties of the two measures of individual reproductive success (i.e.,
 101 IIRS, which is the Intra-population Individual Reproductive Success, and DL), the results obtained
 102 on the DL highlights that the residual variance of SLR can be implicitly explained by annual
 103 environmental variation. The model decomposing the variance in DL measured a larger effect of the
 104 habitat, which is likely due to the fact that DL has different statistical properties. DL accounts for

105 reproductive success, survival and population size. DL is expected to smooth out changes in
106 demographic population growth and to account for the individuals that did not produce offspring
107 but contributed to the population stability or growth by their own survival. Recruitment and
108 population size vary through time in this population (Salles *et al.* 2015), which might explain why
109 more variance in DL was explained by the model. This difference in outcomes among these two
110 estimators of the individual contribution to the local replenishment of the population supports the
111 hypothesis that a part of the self-recruitment variation explained by the habitat is confounded with
112 temporal heterogeneity in population growth and survival.

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##### Heritability Local Fitness Data Clownfish Kimbe Island #####
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#####
##### PACKAGES #####
```

```
library(tnet)
library(pedantics)
library(ggplot2)
library(wesanderson)
library(scales)
library(RColorBrewer)
library(gplots)
library(shape)
library(proxy)
library(MCMCglmm)
```

```
library(nadiv)
library(igraph)
library(lme4)
library(lmerTest)
library(arm)
library(sp)
library(spdep)
library(splm)
library(ape)
library(vegan)
library(ade4)
```

```
library(Cairo)
library(geoR)
library(proxy)
library(pedigree)
```

```
library(FactoMineR)
library(MasterBayes)
library(QGglmm)
library(HDInterval)
```

```
#####
##### PEDIGREE + INVERSE #####
```

```
Ped=as.data.frame(read.table(file="D:/Documents/POSTDOC_TOULOUSE/Publications/T
ransmissionNiche/Data/PED_CLOWNFISH_FINAL_sansNR.txt",header=TRUE,
na.string="NA"))
```

```

Ped$ID=as.factor(Ped$ID)
Ped$FATHER=as.factor(Ped$FATHER)
Ped$MOTHER=as.factor(Ped$MOTHER)

#Orders the pedigree so that parents come before their offspring

library(MasterBayes)
Ped=orderPed(Ped)
#possibility to order the pedigree using fixPedigree if orderPed doesnt work...
Ped=fixPedigree(Ped)
for(x in 1:3) Ped[,x]<-as.factor(Ped[,x])
write.table(Ped,
"D:/Documents/POSTDOC_TOULOUSE/Publications/TransmissionNiche/Data/OrderPed.txt
", sep="\t")

#we keep only informative individuals from the pedigree
Ped=prunePed(Ped, Ped$ID)

#Gets the inverse of the relatedness matrix
library(pedantics)
Ainv=inverseA(Ped)$Ainv
drawPedigree(Ped)

#draw the matrix of genetic similarity
library(nadiv)
pedigree=prepPed(Ped)

#save the figure

jpeg(filename="Models/Figs/Pedigreeall.jpg",
      width=900, height=1000,units="px")
drawPedigree(Ped)
dev.off()

#####
##### DATA REPEATED MEASURES #####

setwd("D:/Documents/POSTDOC_TOULOUSE/Publications/TransmissionNiche/Data")

Aperc=as.data.frame(read.table(file="D:/Documents/POSTDOC_TOULOUSE/Publications
/TransmissionNiche/Data/DATA_CLOWNFISH_FINAL_sansNR_REPEATED_MEASURES.txt",head
er=TRUE,na.string="NA"))

#gets ID/dam/sire as factors
Aperc$ANIMAL=as.factor(Aperc$ID)
Aperc$MOTHER=as.factor(Aperc$MOTHER)
Aperc$FATHER=as.factor(Aperc$FATHER)

#these need to be repeated for connection to the extra matrices in the models
Aperc$ANIMAL2=as.factor(Aperc$ID)

```

```

Aperc$animal=as.factor(Aperc$ID)
Aperc$MOTHER2=as.factor(Aperc$MOTHER)
Aperc$FATHER2=as.factor(Aperc$FATHER)

#these need to be repeated for connection to the extra matrices in the models
Aperc$ANIMAL2=as.factor(Aperc$ID)
Aperc$MOTHER2=as.factor(Aperc$MOTHER)
Aperc$FATHER2=as.factor(Aperc$FATHER)
Aperc$HABITATVIE=as.factor(Aperc$HABITATVIE)
Aperc$HABITATMOTHER=as.factor(Aperc$HABITATMOTHER)

# gets factor or numeric for each parameters
Aperc$sex=as.factor(Aperc$sex)
Aperc$cohort=as.factor(Aperc$cohort)
Aperc$SuccessReproAR=as.numeric(Aperc$SuccessReproAR)
Aperc$Survival=as.numeric(Aperc$Survival)
Aperc$LRS=as.numeric(Aperc$LRS)
Aperc$occurrence=as.numeric(Aperc$occurrence)
Aperc$HabitatBirth=as.numeric(Aperc$HabitatBirth)
Aperc$HabitatResident=as.numeric(Aperc$HabitatResident)
Aperc$LRSbin=as.numeric(Aperc$LRS)
Aperc$LRSbin[Aperc$LRSbin>0]<-1

Aperc$delifing=as.numeric(Aperc$delifing)

Aperc$DepthAnemone=as.numeric(Aperc$DepthAnemone)
Aperc$CodespAnemone=as.factor(Aperc$CodespAnemone)
Aperc$CodeArea=as.factor(Aperc$CodeArea)
Aperc$PropDeadCorals=as.numeric(Aperc$PropDeadCorals)
Aperc$PropCorals=as.numeric(Aperc$PropCorals)
Aperc$PropSand=as.numeric(Aperc$PropSand)
Aperc$PropRock=as.numeric(Aperc$PropRock)
Aperc$PropAlgae=as.numeric(Aperc$PropAlgae)

Aperc$DepthAnemoneParents=as.numeric(Aperc$DepthAnemoneParents)
Aperc$CodeAreaParents=as.factor(Aperc$CodeAreaParents)
Aperc$CodespAnemoneParents=as.factor(Aperc$CodespAnemoneParents)
Aperc$PropDeadCoralsParents=as.numeric(Aperc$PropDeadCoralsParents)
Aperc$PropCoralsParents=as.numeric(Aperc$PropCoralsParents)
Aperc$PropSandParents=as.numeric(Aperc$PropSandParents)
Aperc$PropRockParents=as.numeric(Aperc$PropRockParents)
Aperc$PropAlgaeParents=as.numeric(Aperc$PropAlgaeParents)

#Visualisation des données + sauvegarde
setwd("D:/Documents/POSTDOC_TOULOUSE/Publications/TransmissionNiche/Data")
jpeg("Models/Figs/delifingRM.jpg")
hist(Aperc$delifing)
dev.off()

jpeg("Models/Figs/histLRSRM.jpg")
hist(Aperc$LRS)
dev.off()

```

```
jpeg("Models/Figs/LRSbinRM.jpg")
hist(Aperc$LRSbin)
dev.off()
```

```
#####
##### MODELS #####
```

```
#####"
#### PCI PAPER
```

```
library(MCMCglmm)
```

```
prior5=list(G=list(G1=list(V=1, nu=0.02),
                   G2=list(V=1, nu=0.02),
                   G3=list(V=1, nu=0.02),
                   G4=list(V=1, nu=0.02),
                   G5=list(V=1, nu=0.02)),
            R=list(V=1, nu=0.02))
```

```
MCMCglmm_LRS_pnas5<-MCMCglmm(LRS ~ 1
                             , random=~ANIMAL + MOTHER + HabitatBirth +
HabitatResident + ID
                             , family="poisson"
                             , data= Aperc
                             , ginverse=list(ANIMAL=Ainv)
                             , prior=prior5
                             , nitt = 1000000
                             , burnin = 1000
                             , thin = 1000)
```

```
save(MCMCglmm_LRS_pnas5,
file=paste("Models/", "Figs/", "MCMCglmm_LRS_pnas5.Rdata", sep=""))
```

```
prior4=list(G=list(G1=list(V=1, nu=0.02),
                   G2=list(V=1, nu=0.02),
                   G3=list(V=1, nu=0.02),
                   G4=list(V=1, nu=0.02)),
            R=list(V=1, nu=0.02))
```

```
MCMCglmm_LRS_pnas4<-MCMCglmm(LRS ~ 1
                             , random=~ANIMAL + MOTHER + HabitatBirth + ID
                             , family="poisson"
                             , data= Aperc
                             , ginverse=list(ANIMAL=Ainv)
                             , prior=prior4
                             , nitt = 1000000
                             , burnin = 1000
                             , thin = 1000)
```

```
save(MCMCglmm_LRS_pnas4,
file=paste("Models/", "Figs/", "MCMCglmm_LRS_pnas4.Rdata", sep=""))
```

```
MCMCglmm_LRS_pnas4b<-MCMCglmm(LRS ~ 1
```

```

, random=~ANIMAL + MOTHER + HabitatResident + ID
, family="poisson"
, data= Aperc
, ginverse=list(ANIMAL=Ainv)
, prior=prior4
, nitt = 1000000
, burnin = 1000
, thin = 1000)

save(MCMCglmm_LRS_pnas4b,
file=paste("Models/", "Figs/", "MCMCglmm_LRS_pnas4b.Rdata", sep=""))

MCMCglmm_LRS_pnas4c<-MCMCglmm(LRS ~ 1
, random=~ANIMAL + HabitatBirth + HabitatResident
+ ID
, family="poisson"
, data= Aperc
, ginverse=list(ANIMAL=Ainv)
, prior=prior4
, nitt = 1000000
, burnin = 1000
, thin = 1000)

save(MCMCglmm_LRS_pnas4c,
file=paste("Models/", "Figs/", "MCMCglmm_LRS_pnas4c.Rdata", sep=""))

MCMCglmm_LRS_pnas4d<-MCMCglmm(LRS ~ 1
, random=~MOTHER + HabitatBirth + HabitatResident
+ ID
, family="poisson"
, data= Aperc
, ginverse=list(ANIMAL=Ainv)
, prior=prior4
, nitt = 1000000
, burnin = 1000
, thin = 1000)

save(MCMCglmm_LRS_pnas4d,
file=paste("Models/", "Figs/", "MCMCglmm_LRS_pnas4d.Rdata", sep=""))

#####

MCMCglmm_delifing_pnas5<-MCMCglmm(delifing ~ 1
, random=~ANIMAL + MOTHER + HabitatBirth +
HabitatResident + ID
, family="gaussian"
, data= Aperc
, ginverse=list(ANIMAL=Ainv)
, prior=prior5
, nitt = 1000000
, burnin = 1000
, thin = 1000)

save(MCMCglmm_delifing_pnas5,
file=paste("Models/", "Figs/", "MCMCglmm_delifing_pnas5.Rdata", sep=""))

prior4=list(G=list(G1=list(V=1, nu=0.02),

```

```
G2=list(V=1, nu=0.02),
G3=list(V=1, nu=0.02),
G4=list(V=1, nu=0.02)),
R=list(V=1, nu=0.02))
```

```
MCMCglmm_delifing_pnas4<-MCMCglmm(delifing ~ 1
, random=~ANIMAL + MOTHER + HabitatBirth + ID
, family="gaussian"
, data= Aperc
, ginverse=list(ANIMAL=Ainv)
, prior=prior4
, nitt = 1000000
, burnin = 1000
, thin = 1000)
save(MCMCglmm_delifing_pnas4,
file=paste("Models/", "Figs/", "MCMCglmm_delifing_pnas4.Rdata", sep=""))
```

```
MCMCglmm_delifing_pnas4b<-MCMCglmm(delifing ~ 1
, random=~ANIMAL + MOTHER + HabitatResident + ID
, family="gaussian"
, data= Aperc
, ginverse=list(ANIMAL=Ainv)
, prior=prior4
, nitt = 1000000
, burnin = 1000
, thin = 1000)
save(MCMCglmm_delifing_pnas4b,
file=paste("Models/", "Figs/", "MCMCglmm_delifing_pnas4b.Rdata", sep=""))
```

```
MCMCglmm_delifing_pnas4c<-MCMCglmm(delifing ~ 1
, random=~ANIMAL + HabitatBirth + HabitatResident
+ ID
, family="gaussian"
, data= Aperc
, ginverse=list(ANIMAL=Ainv)
, prior=prior4
, nitt = 1000000
, burnin = 1000
, thin = 1000)
save(MCMCglmm_delifing_pnas4c,
file=paste("Models/", "Figs/", "MCMCglmm_delifing_pnas4c.Rdata", sep=""))
```

```
MCMCglmm_delifing_pnas4d<-MCMCglmm(delifing ~ 1
, random=~MOTHER + HabitatBirth + HabitatResident
+ ID
, family="gaussian"
, data= Aperc
, ginverse=list(ANIMAL=Ainv)
, prior=prior4
, nitt = 1000000
, burnin = 1000
, thin = 1000)
```

```
save(MCMCglmm_delifing_pnas4d,  
file=paste("Models/", "Figs/", "MCMCglmm_delifing_pnas4d.Rdata", sep=""))
```

```
setwd("D:/Documents/POSTDOC_TOULOUSE/Publications/TransmissionNiche/Data/Models  
/Figs")
```

```
load("MCMCglmm_delifing_pnas5.Rdata")  
load("MCMCglmm_delifing_pnas4.Rdata")  
load("MCMCglmm_delifing_pnas4.Rdata")  
load("MCMCglmm_delifing_pnas4b.Rdata")  
load("MCMCglmm_delifing_pnas4c.Rdata")  
load("MCMCglmm_delifing_pnas4d.Rdata")
```

```
load("MCMCglmm_LRS_pnas5.Rdata")  
load("MCMCglmm_LRS_pnas4.Rdata")  
load("MCMCglmm_LRS_pnas4.Rdata")  
load("MCMCglmm_LRS_pnas4b.Rdata")  
load("MCMCglmm_LRS_pnas4c.Rdata")  
load("MCMCglmm_LRS_pnas4d.Rdata")
```

```
summary(MCMCglmm_delifing_pnas5)  
posterior.mode(MCMCglmm_delifing_pnas5$VCV)  
HPDinterval(MCMCglmm_delifing_pnas5$VCV)
```

```
summary(MCMCglmm_delifing_pnas4)  
posterior.mode(MCMCglmm_delifing_pnas4$VCV)  
HPDinterval(MCMCglmm_delifing_pnas4$VCV)  
VA_delifing4= MCMCglmm_delifing_pnas4$VCV[, "ANIMAL"]  
VM_delifing4=MCMCglmm_delifing_pnas4$VCV[, "MOTHER"]  
VNH_delifing4=MCMCglmm_delifing_pnas4$VCV[, "HabitatBirth"]  
VPE_delifing4=MCMCglmm_delifing_pnas4$VCV[, "ID"]  
VR_delifing4=MCMCglmm_delifing_pnas4$VCV[, "units"]
```

```
#heritability  
H2_delifing4=VA_delifing4/(VA_delifing4 + VM_delifing4 + VNH_delifing4 +  
VPE_delifing4 + VR_delifing4)  
posterior.mode(H2_delifing4)  
HPDinterval(H2_delifing4, 0.95)
```

```
summary(MCMCglmm_delifing_pnas4)
```

```
summary(MCMCglmm_delifing_pnas4b)  
posterior.mode(MCMCglmm_delifing_pnas4b$VCV)  
HPDinterval(MCMCglmm_delifing_pnas4b$VCV)  
VA_delifing4b= MCMCglmm_delifing_pnas4b$VCV[, "ANIMAL"]  
VM_delifing4b=MCMCglmm_delifing_pnas4b$VCV[, "MOTHER"]  
VNH_delifing4b=MCMCglmm_delifing_pnas4b$VCV[, "HabitatBirth"]  
VRH_delifing4b=MCMCglmm_delifing_pnas4b$VCV[, "HabitatResident"]
```

```
VPE_delifing4b=MCMCglmm_delifing_pnas4b$VCV[,"ID"]
VR_delifing4b=MCMCglmm_delifing_pnas4b$VCV[,"units"]
```

```
#heritability
```

```
H2_delifing4b=VA_delifing4b/(VA_delifing4b + VM_delifing4b + VRH_delifing4b +
VPE_delifing4b + VR_delifing4b)
posterior.mode(H2_delifing4b)
HPDinterval(H2_delifing4b,0.95)
```

```
summary(MCMCglmm_delifing_pnas4c)
posterior.mode(MCMCglmm_delifing_pnas4c$VCV)
HPDinterval(MCMCglmm_delifing_pnas4c$VCV)
VA_delifing4c= MCMCglmm_delifing_pnas4c$VCV[,"ANIMAL"]
VM_delifing4c=MCMCglmm_delifing_pnas4c$VCV[,"MOTHER"]
VNH_delifing4c=MCMCglmm_delifing_pnas4c$VCV[,"HabitatBirth"]
VRH_delifing4c=MCMCglmm_delifing_pnas4c$VCV[,"HabitatResident"]
VPE_delifing4c=MCMCglmm_delifing_pnas4c$VCV[,"ID"]
VR_delifing4c=MCMCglmm_delifing_pnas4c$VCV[,"units"]
```

```
#heritability
```

```
H2_delifing4c=VA_delifing4c/(VA_delifing4c + VNH_delifing4c + VRH_delifing4c+
VPE_delifing4c + VR_delifing4c)
posterior.mode(H2_delifing4c)
HPDinterval(H2_delifing4c,0.95)
```

```
summary(MCMCglmm_delifing_pnas4d)
posterior.mode(MCMCglmm_delifing_pnas4d$VCV)
HPDinterval(MCMCglmm_delifing_pnas4d$VCV)
```

```
summary(MCMCglmm_LRS_pnas5)
posterior.mode(MCMCglmm_LRS_pnas5$VCV)
HPDinterval(MCMCglmm_LRS_pnas5$VCV)
```

```
summary(MCMCglmm_LRS_pnas4)
posterior.mode(MCMCglmm_LRS_pnas4$VCV)
HPDinterval(MCMCglmm_LRS_pnas4$VCV)
```

```
summary(MCMCglmm_LRS_pnas4b)
posterior.mode(MCMCglmm_LRS_pnas4b$VCV)
HPDinterval(MCMCglmm_LRS_pnas4b$VCV)
```

```
summary(MCMCglmm_LRS_pnas4c)
posterior.mode(MCMCglmm_LRS_pnas4c$VCV)
HPDinterval(MCMCglmm_LRS_pnas4c$VCV)
```

```
summary(MCMCglmm_LRS_pnas4d)
posterior.mode(MCMCglmm_LRS_pnas4d$VCV)
HPDinterval(MCMCglmm_LRS_pnas4d$VCV)
```

```

load("MCMCglmm_delifing_pnas5.Rdata")

summary(MCMCglmm_delifing_pnas5)
VA_delifing=MCMCglmm_delifing_pnas5$VCV[,"ANIMAL"]
mean(VA_delifing)

VM_delifing=MCMCglmm_delifing_pnas5$VCV[,"MOTHER"]
mean(VM_delifing)

VNH_delifing=MCMCglmm_delifing_pnas5$VCV[,"HabitatBirth"]
mean(VNH_delifing)

VRH_delifing=MCMCglmm_delifing_pnas5$VCV[,"HabitatResident"]
mean(VRH_delifing)

VPE_delifing=MCMCglmm_delifing_pnas5$VCV[,"ID"]
mean(VPE_delifing)

VR_delifing=MCMCglmm_delifing_pnas5$VCV[,"units"]
mean(VR_delifing)

posterior.mode(MCMCglmm_delifing_pnas5$VCV)
HPDinterval(MCMCglmm_delifing_pnas5$VCV)
mean(MCMCglmm_delifing_pnas5$VCV)

#heritability
H2_delifing=VA_delifing/(VA_delifing + VM_delifing + VNH_delifing +
VRH_delifing + VPE_delifing + VR_delifing)
mean(H2_delifing)
posterior.mode(H2_delifing)
HPDinterval(H2_delifing,0.95)

#mother effects
M2_delifing=VM_delifing/(VA_delifing + VM_delifing + VNH_delifing +
VRH_delifing + VPE_delifing + VR_delifing)
posterior.mode(M2_delifing)
HPDinterval(M2_delifing,0.95)
mean(M2_delifing)

#evolvability
mdelifing=mean(Aperc$delifing)
Ia_delifing=VA_delifing/(mdelifing*mdelifing)

posterior.mode(Ia_delifing)
HPDinterval(Ia_delifing,0.95)

QGparams(mu=mu,var.a=va,var.p=vp,model="Poisson.log") #regarder si mean.obs et
var.obs sont proches des données obs

mean(Aperc$LRS)

```

```

var(Aperc$LRS)

summary(MCMCglmm_delifing_pnas5)

# model complets MCMCglmm_LRS_pnas5 et MCMCglmm_delifing_pnas5
#Diagnostic of the MCMC

setwd("D:/Documents/POSTDOC_TOULOUSE/Publications/TransmissionNiche/Data")
jpeg("Models/Figs/MCMCglmm_LRS_PLOTVCV.jpg")
plot(MCMCglmm_LRS_pnas5$VCV)
dev.off()

jpeg("Models/Figs/MCMCglmm_delifing_PLOTVCV.jpg")
plot(MCMCglmm_delifing_pnas5$VCV)
dev.off()

#Results from the MCMCglmm
posterior.mode(MCMCglmm_LRS_pnas5$VCV)
HPDinterval(MCMCglmm_LRS_pnas5$VCV)

summary(MCMCglmm_LRS_pnas5)

VA_LRS= MCMCglmm_LRS_pnas5$VCV[, "ANIMAL"]
VM_LRS=MCMCglmm_LRS_pnas5$VCV[, "MOTHER"]
VNH_LRS=MCMCglmm_LRS_pnas5$VCV[, "HabitatBirth"]
VRH_LRS=MCMCglmm_LRS_pnas5$VCV[, "HabitatResident"]
VPE_LRS=MCMCglmm_LRS_pnas5$VCV[, "ID"]
VR_LRS=MCMCglmm_LRS_pnas5$VCV[, "units"]

mean(VA_LRS)
mean(VM_LRS)
mean(VNH_LRS)
mean(VRH_LRS)
mean(VPE_LRS)
mean(VR_LRS)

#heritability
H2_LRS=VA_LRS/(VA_LRS + VM_LRS + VNH_LRS + VRH_LRS + VPE_LRS + VR_LRS)
mean(H2_LRS)
posterior.mode(H2_LRS)
HPDinterval(H2_LRS,0.95)

#mother effects
M2_LRS=VM_LRS/(VA_LRS + VM_LRS + VNH_LRS + VRH_LRS + VPE_LRS + VR_LRS)
mean(M2_LRS)
posterior.mode(M2_LRS)
HPDinterval(M2_LRS,0.95)

#evolvability equation:  $I_a=V_a/m^2$  with m the mean of the trait
mLRS=mean(Aperc$LRS)
Ia_LRS=mean(VA_LRS)/(mLRS*mLRS)
mean(Ia_LRS)
posterior.mode(Ia_LRS)
HPDinterval(Ia_LRS,0.95)

```

```

library(QGglmm)
library(MCMCglmm)
#heritability estimate using de Villmereuille et al 2016 Genetics
mu= mean(MCMCglmm_LRS_pnas5[["Sol"]][, "(Intercept)"])
va= mean(MCMCglmm_LRS_pnas5[["VCV"]][, "ANIMAL"])
vm=mean(MCMCglmm_LRS_pnas5[["VCV"]][, "MOTHER"])
vnh=mean(MCMCglmm_LRS_pnas5[["VCV"]][, "HabitatBirth"])
vrh=mean(MCMCglmm_LRS_pnas5[["VCV"]][, "HabitatResident"])
vpe=mean(MCMCglmm_LRS_pnas5[["VCV"]][, "ID"])
vr=mean(MCMCglmm_LRS_pnas5[["VCV"]][, "units"])
vp=mean(rowSums(MCMCglmm_LRS_pnas5[["VCV"]]))

QGparams(mu=mu, var.a=va, var.p=vp, model="Poisson.log") #héritabilité
QGparams(mu=mu, var.a=vm, var.p=vp, model="Poisson.log") # effets maternels
QGparams(mu=mu, var.a=vnh, var.p=vp, model="Poisson.log") #HabitatBirth
QGparams(mu=mu, var.a=vrh, var.p=vp, model="Poisson.log") #HabitatResident
QGparams(mu=mu, var.a=vpe, var.p=vp, model="Poisson.log") #ID
QGparams(mu=mu, var.a=vr, var.p=vp, model="Poisson.log") #units

QGicc(mu=mu, var.comp=vm, var.p=vp, model="Poisson.log") # effets maternels
QGicc(mu=mu, var.comp=vnh, var.p=vp, model="Poisson.log") #HabitatBirth
QGicc(mu=mu, var.comp=vrh, var.p=vp, model="Poisson.log") #HabitatResident
QGicc(mu=mu, var.comp=vpe, var.p=vp, model="Poisson.log") #ID
QGicc(mu=mu, var.comp=vr, var.p=vp, model="Poisson.log") #units

QGicc(mu = mu, var.p = vp, var.comp = va, model = "Poisson.log")#Computing the
broad - sense heritability
QGicc(mu = mu, var.p = vp, var.comp = vm, model = "Poisson.log")

#récupérer les posteriors distributions

df= data.frame(mu = as.vector(MCMCglmm_LRS_pnas5[["Sol"]][, "(Intercept)"]),
va = as.vector(MCMCglmm_LRS_pnas5[["VCV"]][, "ANIMAL"]),
vm= as.vector(MCMCglmm_LRS_pnas5[["VCV"]][, "MOTHER"]),
vnh=mean(MCMCglmm_LRS_pnas5[["VCV"]][, "HabitatBirth"]),
vrh=mean(MCMCglmm_LRS_pnas5[["VCV"]][, "HabitatResident"]),
vpe=mean(MCMCglmm_LRS_pnas5[["VCV"]][, "ID"]),
vr=mean(MCMCglmm_LRS_pnas5[["VCV"]][, "units"]),
vp = rowSums(MCMCglmm_LRS_pnas5[["VCV"]]))

head(df)

# on calcule les posterior distribution pour h²
post= do.call("rbind", apply(df, 1, function(row){
QGparams(mu=row[["mu"]],
var.a=row[["va"]],
var.p=row[["vp"]],
model="Poisson.log", verbose= FALSE)
}))

head(post) # on regarde la tête du tableau

```

```

post.h2.obs=as.matrix(post$h2.obs) # on sélectionne la colonne h2
mean(post.h2.obs)
post.var.a.obs=as.matrix(post$var.a.obs)
hdi(post.var.a.obs, credMass = 0.95)# on mesure la CI95%
hdi(post.h2.obs, credMass = 0.95)

#estimation de l'évolvabilité du trait LRS
post_Ia_LRS_obervedscale= post$var.a.obs / (mLRS*mLRS)
mean(post_Ia_LRS_obervedscale)
posterior.mode(post_Ia_LRS_obervedscale)
hdi(post_Ia_LRS_obervedscale,0.95)

# on calcule les posterior distribution pour m2
postm= do.call("rbind", apply(df, 1, function(row){
  QGparams(mu=row[["mu"]],
           var.a=row[["vm"]],
           var.p=row[["vp"]],
           model="Poisson.log", verbose= FALSE)
}))

head(postm) # on regarde la tête du tableau
postm.h2.obs=as.matrix(postm$h2.obs) # on sélectionne la colonne h2
postm.var.a.obs=as.matrix(postm$var.a.obs)
hdi(postm.h2.obs, credMass = 0.95)# on mesure la CI95%
hdi(postm.var.a.obs, credMass = 0.95)

# on calcule les posterior distribution pour m2 avec QGicc
postm= do.call("rbind", apply(df, 1, function(row){
  QGicc(mu=row[["mu"]],
        var.comp=row[["vm"]],
        var.p=row[["vp"]],
        model="Poisson.log", verbose= FALSE)
}))

head(postm) # on regarde la tête du tableau
postm.var.comp.obs=as.matrix(postm$var.comp.obs)
hdi(postm.var.comp.obs, credMass = 0.95)

postm.m2.obs=as.matrix(postm$icc.obs) # on sélectionne la colonne h2
hdi(postm.m2.obs, credMass = 0.95)# on mesure la CI95%

# on calcule les posterior distribution pour vnh
postvnh= do.call("rbind", apply(df, 1, function(row){
  QGicc(mu=row[["mu"]],
        var.comp=row[["vnh"]],
        var.p=row[["vp"]],

```

```

        model="Poisson.log", verbose= FALSE)
    )))
head(postvnh) # on regarde la tête du tableau
postvnh.var.comp.obs=as.matrix(postvnh$var.comp.obs) # on sélectionne la
colonne h2
hdi(postvnh.var.comp.obs, credMass = 0.95)# on mesure la CI95%

# on calcule les posterior distribution pour vrh
postvrh= do.call("rbind", apply(df, 1, function(row){
  QGicc(mu=row[["mu"]],
        var.comp=row[["vrh"]],
        var.p=row[["vp"]],
        model="Poisson.log", verbose= FALSE)
}))
head(postvrh) # on regarde la tête du tableau
postvrh.var.comp.obs=as.matrix(postvrh$var.comp.obs) # on sélectionne la
colonne h2
hdi(postvrh.var.comp.obs, credMass = 0.95)# on mesure la CI95%

# on calcule les posterior distribution pour vpe
postvpe= do.call("rbind", apply(df, 1, function(row){
  QGicc(mu=row[["mu"]],
        var.comp=row[["vpe"]],
        var.p=row[["vp"]],
        model="Poisson.log", verbose= FALSE)
}))
head(postvpe) # on regarde la tête du tableau
postvpe.var.comp.obs=as.matrix(postvpe$var.comp.obs) # on sélectionne la
colonne h2
hdi(postvpe.var.comp.obs, credMass = 0.95)# on mesure la CI95%

# on calcule les posterior distribution pour vr
postvr= do.call("rbind", apply(df, 1, function(row){
  QGicc(mu=row[["mu"]],
        var.comp=row[["vr"]],
        var.p=row[["vp"]],
        model="Poisson.log", verbose= FALSE)
}))
head(postvr) # on regarde la tête du tableau
postvr.var.comp.obs=as.matrix(postvr$var.comp.obs) # on sélectionne la colonne
h2
hdi(postvr.var.comp.obs, credMass = 0.95)# on mesure la CI95%

posterior.mode(MCMCglmm_LRS_pnas5$VCV)#estimation of additive genetics ans

```

```
residual variance
HPDinterval(MCMCglmm_LRS_pnas5$VCV) #interval de confiance (95%) du mod
distribution.
```

```
##### Utiliser QGparams pour vérifier que Poisson log est bien pour LRS
# QGparams renvoie moyenne et variance des données qui correspondent à
l'estimation du modèle
```

```
setwd("D:/Documents/POSTDOC_TOULOUSE/Publications/TransmissionNiche/Data/Models
/Figs")
load("MCMCglmm_LRS_pnas5.Rdata")
QGparams(mu=mu,var.a=va,var.p=vp,model="Poisson.log") #regarder si mean.obs et
var.obs sont proches des données obs
```

```
mean(Aperc$LRS)
var(Aperc$LRS)
```

```
summary(MCMCglmm_delifing_pnas5)
setwd("D:/Documents/POSTDOC_TOULOUSE/Publications/TransmissionNiche/Data/Models
/Figs")
```

```
#####
###
```